

Cloud Point Extraction of Carbendazim Pesticide in Foods and Environmental Matrices Prior to Visible Spectrophotometric Determination

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Abstract: Two simple eco-friendly methods are described for nano-determination of carbendazim (MBC) pesticide in real samples. These methods are based on oxidation of MBC pesticide with Fe (III) ions in acidic medium. The formed Fe(II) ions reacts with potassium ferricyanide to form blue colored product (method A) which can easily be extracted into nonionic surfactant solution of Triton X-114 at cloud point temperature (CPT) of 55°C and MBC determined spectrophotometrically at absorption maximum of 685 nm with apparent molar absorptivity of $2.07 \times 10^4 \text{ L mol}^{-1} \text{ cm}^{-1}$. The Method B is based on the reaction of the formed Fe (II) with 2, 2'-bipyridyl to form a stable orange colored complex which can also be extracted by Triton X-114 at the same CPT and MBC determined spectrophotometrically at absorption maximum of 521 nm with apparent molar absorptivity of $1.83 \times 10^4 \text{ L mol}^{-1} \text{ cm}^{-1}$. Optimization of the experimental parameters was described and interferences study also examined. Under the optimum conditions established, the calibration graphs for MBC were linear in the range of 0.5-13 and 1-20 ng mL⁻¹, giving the detection limits of 0.46 and 0.49 ng mL⁻¹ with enrichment factors of 85.7 and 38.9 fold for method A and B respectively. The average percent recoveries in the real spiked samples were (97.86±1.06%) and (98.66±0.93%), giving a precision in terms of %RSD in the range of 1.25-2.97% and 0.37-1.42% for method A and B respectively. The proposed methods were applied to the determination of MBC in vegetables, orange, and water samples.

Keywords: Carbendazim, Vegetables and Waters, Cloud Point Extraction, Visible Spectrophotometry

1. Introduction

The use of pesticides is considered a double-edged sword; they are very useful and important in addressing a broad range of diseases caused by various harmful insects in the agricultural products, thereby enhancing food production. But, at the same time it pays inevitably toward pollution of various environmental components and, therefore, is hazardous to human and animal health. In this regard, the massive use and/or misuse of any pesticide may lead to environmental problems and several poisoning cases in human beings and other organisms. Carbendazim (MBC) chemically named by IUPAC as methyl benzimidazol-2-yl carbamate (Figure 1) is a benzimidazolic systemic fungicide widely used in agriculture for controlling several diseases on the fruits, vegetables,

tobacco, cotton and cereals, and also used in post-harvest protection of crops against fungal diseases [1-2]. Carbendazim is degraded slowly in the environment because the nature of benzimidazolic ring in its structure difficult to break thus it persists for a long time in the environment [3].

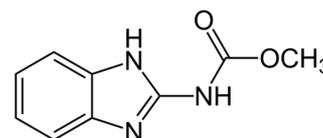


Figure 1. The structural formula of Carbendazim (chemical formula: $C_9H_9N_3O_2$, mol.wt. 191.21 g mol⁻¹).

In fact, this property makes this pesticide a very serious problem for the environment and has the direct detrimental

effect on human and animal health, and on this basis it is listed in toxicity class IV pesticides [4]. Accordingly, different international legislations bodies such as the European Union (EU) [5], United States Environmental Protection Agency (US EPA) [6], World Health Organization (WHO) [7] and other international bodies have been established the maximum residues level (MRL) set for MBC, for example, orange in the EU is set at 0.2 mg/kg, cucumber in the FAO/WHO set at 0.05 mg kg⁻¹, food in China set at 0.5 mg/kg [8] and a Brazilian regulatory agency, sets a limit of 0.02 mg kg⁻¹ as the human acceptable daily intake (ADI) of MBC [9]. In light of this information, importance of continuous controlling and assessment for the presence of MBC residue at low levels in a variety of samples is a must and needs to establish sensitive, selective and reliable methods with high speed.

Several analytical techniques for MBC determination have been reported in chemical literature, but the most common one based on the sophisticated techniques, and on the particular the chromatographic methods including high-performance liquid chromatography (HPLC) [1, 10-12], high-performance liquid chromatography–mass spectrometry (HPLC-MS) [13-15], gas chromatography(GC) [16] and ion chromatography with fluorescence detector [17]. In addition, other techniques were also reported such electro-analytical methods [18-21], immunoassay [22], fluorescence spectrometry [23]. Since these methods have certainly high-sensitivity and low detection limit, but they are expensive, tedious and not available in most laboratories. UV-Visible spectrophotometric technique has a limited use in the determination of MBC due to the lack of its sensitivity and therefore, a scarce papers have been appeared in literature [24]. Because, in general, the concentration of pesticides in different types of food and environmental samples is very low, a pre-concentration step to quantify these compounds is a must. A number of extraction procedures coupled with different instrumental techniques have recently developed for extraction and enrichment of MBC in a variety of samples including solid phase extraction [25-27], dispersive liquid-liquid microextraction [28-31], ionic liquid-dispersive liquid-liquid microextraction [32], cloud point extraction (CPE) is now becoming a well-established and accepted as an alternative method of extraction / enrichment methodology of organic pollutants including pesticides in various matrices due to its simplicity, eco-friendly, cheap and relatively high extraction efficiency [33-37].

In this work, we present two spectrophotometric methods for the determination of MBC after cloud point extraction (CPE) in different samples for the first time. These methods are based on the oxidation of MBC with Fe (III) ions in acidic medium. The formed Fe (II) ions react with potassium ferricyanide to form blue colored product (method A) and with 2, 2'-bipyridyl to form a stable orange colored complex (method B). These two colored products can easily extract into nonionic surfactant solution of Triton X-114 and MBC determined spectrophotometrically at each respective absorption maximum.

2. Materials and Methods

2.1. Apparatus

All absorption spectra and absorbance measurements for MBC throughout this study were carried out by using a Shimadzu double-beam UV-Vis Spectrophotometer model UV-1800 (Kyoto, Japan) working at a wavelength of 190-1100 nm, and equipped with 5-mm optical path cell. For the solution pH measurement, a portable pH-meter microprocessor (HANNA, Germany) was used. The shaking water bath SW23 microprocessor with PID temperature control (JULABO GmbH, Germany) was employed during the course of CPE experiments.

2.2. Reagents and Materials

All materials and reagents used in this work with high purity and doubly distilled water used in the preparation of all solutions and for the final rinsing of glass wares. Carbendazim (MBC) (99.0% purity,) was purchased from Accustandard[®] (Connecticut, USA). A stock solution (100 µg mL⁻¹) of MBC was prepared by dissolving 10 mg in 20 mL of 0.1 N HCl in a 100 mL volumetric flask and diluted to mark with distilled water and kept in an amber bottle in the refrigerator. Triton X-114 (purity >99.9%), was purchased from AMRESCO LLC (Solon, USA). A 10% (v/v) of Triton X-114 was prepared by diluting 10 mL in 100 mL water. A 1x10⁻³ M of FeCl₃ (99.9%, BDH) was prepared by dissolving 0.0162 g in 5 mL water and diluted to mark in 100 mL volumetric flask. A 0.01 M potassium ferricyanide (99.0%, Sigma-Aldrich) was prepared by dissolving 0.3292 g in 5 mL water and diluted to mark in 100 mL volumetric flask. A 1x10⁻³ M of 2, 2-Bipyridyl (99.0%, BDH) was prepared by dissolving 0.0156 g in 5 mL water and diluted to mark in 100 mL volumetric flask. Acetate buffer solutions were prepared from different volumes of 0.1M of acetic acid (>99%, Sigma-Aldrich) and 0.1 M sodium acetate (99.0%, Merck). Orthophosphoric acid (85%, BDH) was prepared by appropriate dilution of concentrated acid with water. Sodium sulphate, sodium acetate and magnesium sulphate 6-hydrates were purchased from Riedel-deHaën AG (Germany). Acetonitrile was obtained from BDH (England). Carbograph and an ion-ion exchange (PSA) were purchased from Sigma-Aldrich (USA) and Vertical Chromatography Co., Ltd. (Thailand) respectively.

2.3. Recommended CPE Procedures

2.3.1. Method A

Aliquots of MBC standard or sample solution ranging from 0.05-1.3 mL of 100 ng mL⁻¹ which corresponding to 0.5-13 ng mL⁻¹ of MBC were transferred into a series of 10 mL centrifugal tubes. To each tube 0.4 mL of 1x10⁻³ M FeCl₃ solution, 0.4 mL of 0.01 M K₃Fe(CN)₆ solution and 1 mL of acetate buffer (pH=4) were added, then kept the solution on water bath at 60°C for 15 min and cooled thereafter. Then, 1.0 mL of 1N H₃PO₄ and 0.8 mL of 10% Triton X-114 were added. The content of each tube was made up to 10 mL with water.

All tubes were transferred into a water bath at 55°C for 15 min to induce the formation of cloudy solution and centrifuged at 3500 rpm for 20 min to separate the two phases. After decantation of the aqueous phase, the surfactant-rich phase that remained adhered to the tube was dissolved with a 1.0 mL of ethanol: water (1:1) and the absorbance of each solution containing MBC was measured spectrophotometrically in 5-mm quartz cell at λ_{\max} of 685 nm against a reagent blank solution.

2.3.2. Method B

Aliquots of MBC standard or sample solution ranging from 0.1-2.0 mL of 100 ng mL⁻¹ which corresponding to 1-20 ng mL⁻¹ MBC were transferred into a series of 10 mL centrifugal tubes. To each tube 0.8 mL of 1x10⁻³ M 2,2-Bipyridyl solution, 0.8 mL of 1x10⁻³ M FeCl₃ solution and 1.0 mL of acetate buffer (pH=3) were added and kept on water bath at 50°C for 15 min and cooled thereafter. Then, 0.6 mL of 10% Triton X-114 was added and the content of each tube diluted to 10 mL. All tubes were transferred into a water bath at 60°C for 30 min to induce the formation of a cloudy solution and centrifuged at 3500 rpm for 20 min to separate the two phases. After decantation of the aqueous phase, the surfactant-rich phase that remained adhered to the tube was dissolved with a 1.0 mL of ethanol: water (1:1) and the absorbance of each solution containing MBC was measured spectrophotometrically in 5-mm quartz cell at λ_{\max} of 521 nm against a reagent blank solution.

2.4. Sample Preparation

2.4.1. Water

About one liter of drinking and river water samples was randomly collected from the campus of University of Baghdad / Iraq. The river water was first filtered off to remove any suspended materials and all samples were kept in the refrigerator until analyzed. Each sample was spiked with different concentration of MBC standard and subjected to recommended CPE procedures (A and B) and MBC was determined by spectrophotometry at λ_{\max} of 685 and 521 nm respectively, from the constructed calibration curves.

2.4.2. Soil

The soil sample was randomly collected from the home garden and the soil sample solution was prepared according to the procedure adopted by Pourreza et al [31] with little modification. The sample was air-dried at room temperature, ground in agate mortar into small particle size of about 250 μ m sieves and stored in a closed vessel. 20 g of sample was weighted in 100 mL conical flasks and 40 mL of 0.1 M HCl was added. The content was shaken in a mechanical shaker for one hr., then filtered and the pH of the filtered was adjusted to 7.0 by diluted NaOH. Three portions of the resultant solution were directly spiked with different concentration of MBC standard solutions and subjected to recommend CPE procedures (A and B) and MBC was determined by spectrophotometry at λ_{\max} of 685 and 521 nm respectively from the constructed calibration curves.

2.4.3. Vegetables and Orange

Vegetables (Cucumber and Tomato) and an orange were purchased from local markets in Baghdad, Iraq. The QuEChERS (quick, easy, cheap, effective, rugged, and safe) method [38] used for pesticide residue analysis was adopted with little modification for sample preparation of vegetables and orange. A 0.5 kg of each sample was selected and the edible part was cut into 1-cm pieces and blended using a commercial food mixer for homogenization of the sample. A 15 g sample portion was placed in 100 mL conical flask and 20 mL of solvent mixture containing acetic acid and acetonitrile (1:5) was added and the content was shaken vigorously in an electrical shaker for one hr. After shaking, the extract was withdrawn and transferred into 50 mL centrifugal tube and mixed with 10 g sodium sulphate, 4 g magnesium sulphate, 1 g of sodium acetate and centrifuged for 10 min at 250 rpm to separate the phases. The upper layer was taken and mixed with 0.3 g PSA and 0.6 g Carbograph in another centrifugal tube and immediately shaken and filtered. The filtrate solution was evaporated at 50°C on water bath to remove the solvent. The residue was dissolved with water then diluted to 10 mL in standard volumetric flask. Each sample solution was spiked with different concentration of MBC standard solutions and subjected to recommend CPE procedures (A and B) and MBC was determined by spectrophotometry at λ_{\max} of 685 and 521 nm respectively from the constructed calibration curves.

2.5. Statistical Analysis

Excel 2007 (Microsoft Office®) and Minitab version 17(Minitab Inc., State College, PA, USA) were employed to carry out all statistical calculations such as regression and correlation analysis, ANOVA and significance tests.

3. Results and Discussion

3.1. Mechanism of Reactions

It is undoubtedly proved that iron (III) salts can act as oxidants for most organic compounds in certain experimental conditions leading to the formation of oxidizing organic product and Fe (II) as reduced form of Fe (III) [39-41]. This idea has been utilized to design two analytical methods for the determination of MBC using the combined CPE-Spectrophotometry. To act as an oxidant, Fe (III) salt can reduce to Fe (II) salt which is equivalent to the amount of organic material [42]. The amount of iron (II) formed can then be determined spectrophotometrically by complexing with the familiar conventional reagents. Thus in method A, Fe(II) ion is formed via the reduction of Fe(III) by MBC pesticide and subsequent reaction with potassium ferricyanide (PFC), forming a precipitate blue product (insoluble bright blue pigment called Turnbull's) which is soluble in acidic medium as shown in the Figure 2.

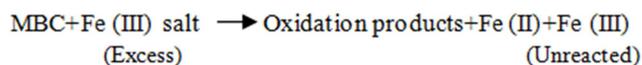


Figure 2. The reaction path of the method A.

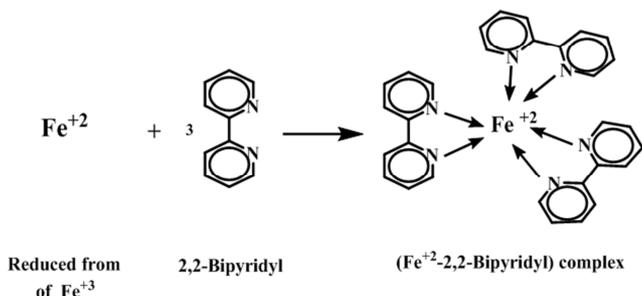
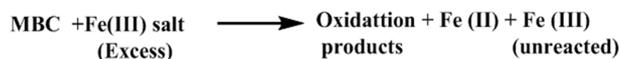


Figure 3. The reaction path of the method B.

Upon addition of H_3PO_4 and buffer solution (pH=3), the

soluble colored product is formed which can be easily extracted into Triton X-114. Preliminary experiments have shown that the absorption maximum of the colored product at different concentration of MBC in micelle-mediating extraction occurs at 685 nm and the absorbance increases linearly with increasing the pesticide concentration. In Method B, the reduced form of Fe III (Fe II) can form a chelate with 2, 2'-bipyridyl (Figure 3) giving a stable orange colored complex exhibit absorption maximum at 521 nm in surfactant-rich phase against the reagent blank.

3.2. Absorption Spectra

The absorption spectra of the two colored products were recorded in the presence of surfactants against a reagent blank prepared under optimum conditions. The spectra of blue colored (method A) and orange product (method B) show the absorption maxima of 658 and 521 nm with molar absorptivities (ϵ) of 2.07×10^4 and $1.83 \times 10^4 \text{ L mol}^{-1} \text{ cm}^{-1}$ obtained respectively, while other reagents gave different absorption maxima in UV region as displayed in Figures 4 and 5. Thus, these absorption maxima of the colored products were adopted throughout this study.

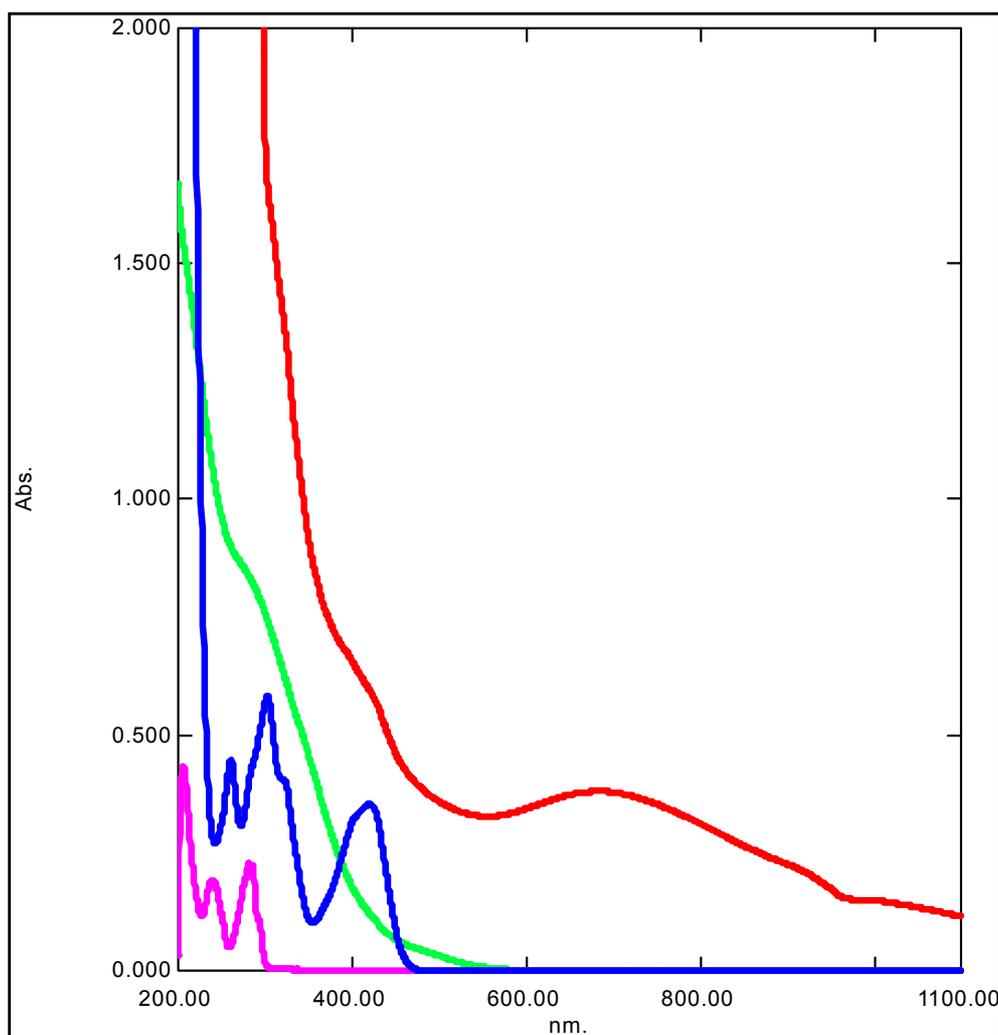


Figure 4. Absorption spectra of (a) Carbendazim solution (green color) (b) potassium ferric cyanide solution (blue color) (c) FeCl_3 solution (pink color) (d) Colored product in surfactant rich-phase (red color).

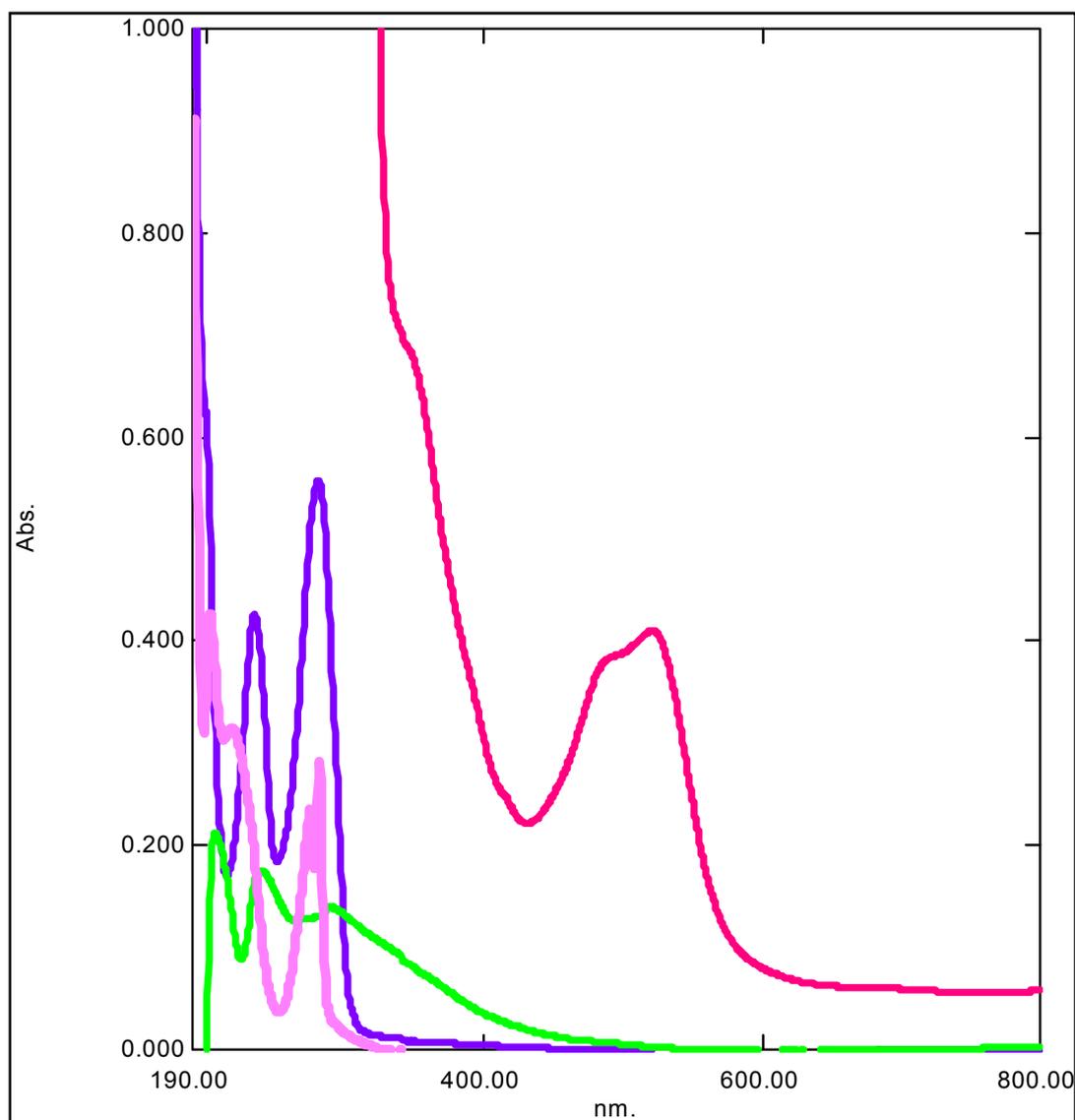


Figure 5. Absorption spectra of (a) Carbendazim solution (green color) (b) (c) 2,2'-bipyridyl (pink color) (c) FeCl_3 solution (purple color) (d) Colored product in surfactant rich-phase (red color).

3.3. Optimization of CPE Procedure

The influence of various parameters such as, pH, FeCl_3 , $\text{K}_3\text{Fe}(\text{CN})_6$ and 2,2'-bipyridyl concentration, H_3PO_4 concentration, Triton X-114 amount, equilibrium temperature and incubation time were investigated in detail by classical optimization to maximize the analytical figures of merit and the extraction efficiency of MBC.

3.3.1. Effect of pH

The solution pH is an important factor affecting the absorbance of the surfactant-rich phase (SRP) and thus the extraction efficiency of MBC for the two methods. Therefore, the CPE was carried out at temperature 70°C for 30 min in the solutions via varying the pH values within 2.0-7.0 at MBC concentration of 5 ng ml^{-1} (method A) and 12 ng ml^{-1} (method B) keeping other parameters such as FeCl_3 , $\text{K}_3\text{Fe}(\text{CN})_6$, H_3PO_4 , 2,2'-bipyridyl and Triton X-114 at concentration of $1 \times 10^{-4} \text{ M}$, $1 \times 10^{-3} \text{ M}$, 0.05 N, $1 \times 10^{-4} \text{ M}$ and 1% in final 10 mL aqueous

solution respectively. The results are depicted in Figure 6. It can be noted that the absorbance at highest at pH of 4.0 for method A and 3.0 for method B, then decreased thereafter. Thus, these pH values were adopted in the further experiments.

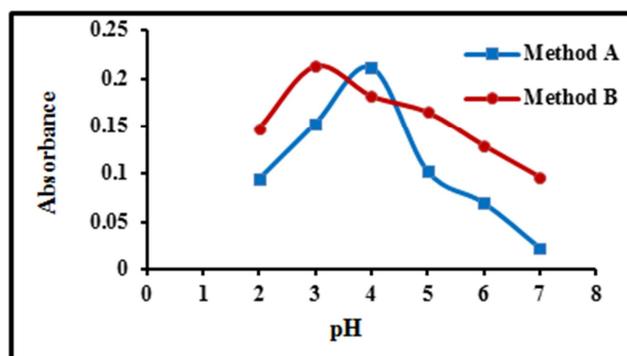


Figure 6. Effect of pH.

3.3.2. Effect of $FeCl_3$ Concentration

The influence of $FeCl_3$ concentration on the formation of blue product and complex formation for the two methods was examined by varying the volume ranging from 0.1 to 1.2 mL of 1×10^{-3} M $FeCl_3$ solution at pH 4.0 (method A) and 3.0 (method B), keeping other parameters constant. The results displayed in Figure 7 revealed that maximum absorbance was achieved when $FeCl_3$ concentration was of 4×10^{-5} M (0.4 mL of 1×10^{-3} M in 10 mL final aqueous solution) for method A and 8×10^{-5} M (0.8 mL of 1×10^{-3} M in 10 mL final aqueous solution) for method B. Consequently, these concentrations were selected as optimal for the next experiments.

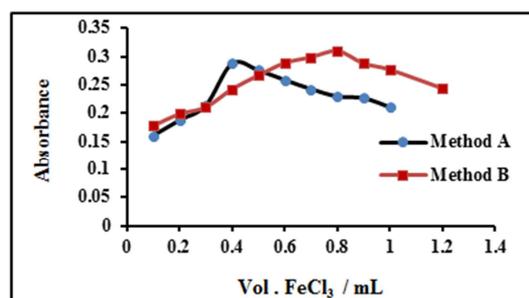


Figure 7. Effect of $FeCl_3$ concentration [Conditions: method A: 5 ng mL^{-1} MBC; $K_3Fe(CN)_6$ concentration, 1×10^{-4} M; pH 4.0; H_3PO_4 , 0.05 N, TX-114, 1%; method B: 12 ng mL^{-1} MBC; 2,2'-bipyridyl, 1×10^{-4} M; pH, 3.0; TX-114, 1%; equilibration temperature and incubation temperature, 70°C for 30 min for two method].

3.3.3. Effect of Reagent Concentration

The effect of $K_3Fe(CN)_6$ and 2,2'-bipyridyl reagents concentration on the absorbance signal of MBC was seeking by varying the volume from 0.1-1.0 mL of 1×10^{-3} M of both reagents. The results shown in Figure 8 appeared that the absorbance was linearly increased with increasing the reagents concentration and reached maximum at 4×10^{-5} M (0.4 mL of 1×10^{-3} in 10 mL solution) $K_3Fe(CN)_6$ and 8×10^{-5} M (0.8 mL of 1×10^{-3} in 10 mL solution) 2,2'-bipyridyl concentration. Therefore, these concentrations found to be a highly suitable for the formation of the resultant colored product and complex and thus it was adopted for further experiments as optimal.

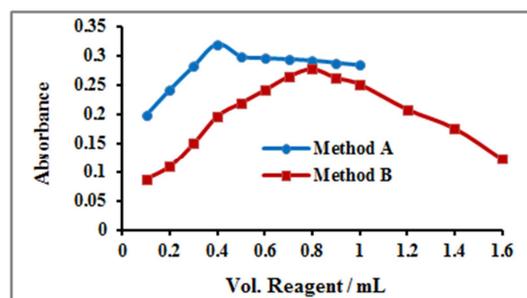


Figure 8. Effect of $K_3Fe(CN)_6$ and 2,2'-bipyridyl reagents concentration [Conditions: method A: 5 ng mL^{-1} MBC; $FeCl_3$ concentration, 4×10^{-5} M; pH 3.0; H_3PO_4 , 0.05N, TX-114, 1%; method B: 12 ng mL^{-1} MBC; $FeCl_3$ concentration, 8×10^{-4} M; pH, 4.0; TX-114, 1%; equilibration temperature and incubation temperature, 70°C for 30 min for two method].

3.3.4. Effect of H_3PO_4 Concentration

It was observed that Turnbull's blue product formed before

extraction in method A is a slightly soluble in the surfactant medium used and the extraction efficiency of MBC was very poor. One of previous studies[31] have shown that the presence phosphoric acid enhances the solubility of the blue product before the extraction process. For this, the effect of different concentrations of H_3PO_4 was conducted by varying the volume from 0.2-2.0 mL of 1.0 N H_3PO_4 on the absorbance of the colored product. It can be seen (Figure 9) that the optimum concentration of H_3PO_4 was of 0.1 N in the final aqueous solution and thus it was selected in following experiments.

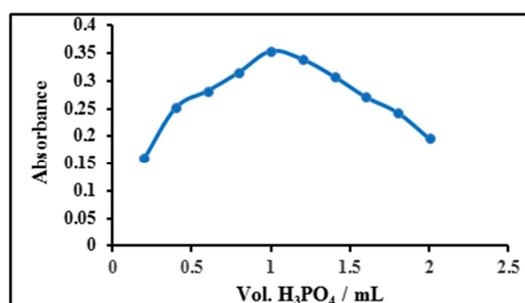


Figure 9. Effect of H_3PO_4 concentration [Conditions: method A: 5 ng mL^{-1} MBC; $FeCl_3$ concentration, 4×10^{-5} M; $K_3Fe(CN)_6$ concentration, 4×10^{-5} M; pH 3.0; TX-114, 1%; equilibration temperature and incubation temperature, 70°C for 30 min].

3.3.5. Effect of Surfactant Amount

The variation of the absorbance with surfactant (Triton X-114) amount on the extraction of MBC pesticide was studied within the volume range 0.1-1.0 mL of 10% Triton X-114. As shown in Figure 10, at a lower amount of surfactant, the absorbance was low for the two methods and maximum remarkable extraction was observed for Triton X-114 amount of 0.8 % for the method A and 0.6% for method B in the final 10 mL aqueous solution which gave the best preconcentration factor. Therefore, these values were adopted in the recommended CPE procedure.

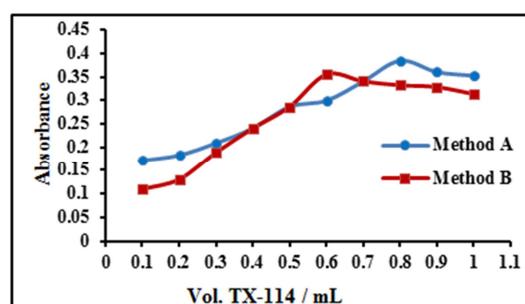


Figure 10. Effect of Triton X-114 amount [Conditions: method A: 5 ng mL^{-1} MBC; $FeCl_3$ concentration, 4×10^{-5} M; $K_3Fe(CN)_6$ concentration, 4×10^{-5} M; pH 3.0; H_3PO_4 , 0.1 N; method B: 12 ng mL^{-1} MBC; $FeCl_3$ concentration, 8×10^{-4} M; 2,2'-bipyridyl concentration, 8×10^{-4} M; pH, 4.0; equilibration temperature and incubation temperature, 70°C for 30 min for two method].

3.3.6. Effect of Temperature and Time

These two parameters play an important role in the CPE process for completion of the reaction and to achieve the best extraction of the target analyte. Thus the effect of equilibrium temperature and incubation time was varied in the range of

20-75°C and 5-50 min respectively as showed in Figures 9 and 10. It was found (Figure 11) that the maximum absorbance was obtained at 55°C for 15 min and 60°C for 30 min for the method A and B respectively, and they were nearly constant above these values. Accordingly, 55°C and 60°C were used in the recommended CPE procedure. The study of incubation time (Figure 12) also indicated that the maximum absorbance value was achieved at 15 and 30 min for the method A and B respectively. Thus, they were used as an optimal in the recommended CPE procedure.

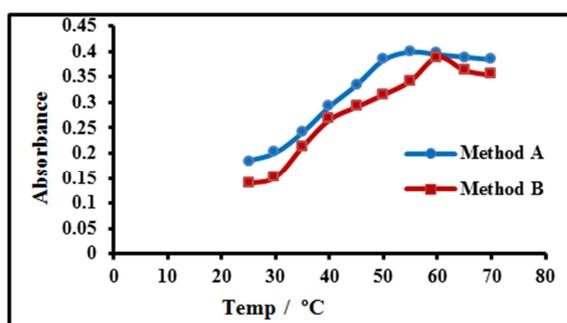


Figure 11. Effect of Temperature [Conditions: method A: 5 ng mL⁻¹ MBC; FeCl₃ concentration, 4x10⁻³ M; K₃Fe(CN)₆ concentration, 4x10⁻⁵ M; pH 3.0; 0.H₃PO₄, 0.1 N; TX-114, 0.8%; method B: 12 ng mL⁻¹ MBC; FeCl₃ concentration, 8x10⁻⁴ M; 2,2'-bipyridyl concentration, 8x10⁻⁴ M; pH, 4.0; TX-114, 0.6%; incubation time, 30 min for two method].

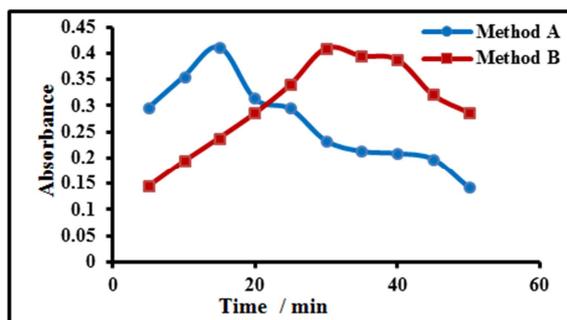


Figure 12. Effect of incubation time [Conditions: method A: 5 ng mL⁻¹ MBC; FeCl₃ concentration, 4x10⁻³ M; K₃Fe(CN)₆ concentration, 4x10⁻⁵ M; pH 3.0; 0.H₃PO₄, 0.1 N; TX-114, 0.8%; method B: 12 ng mL⁻¹ MBC; FeCl₃ concentration, 8x10⁻⁴ M; 2,2'-bipyridyl concentration, 8x10⁻⁴ M; pH, 4.0; TX-114, 0.6%; equilibration temperature, 55°C and 60°C for method A and B respectively].

3.4. Order of Additions

The effect of order for additions on the absorption signal of the colored product and the complex was also investigated. Table 1 reveals that the best order of addition was number 2 and 3 for the method A and B respectively, due to give a highest absorption signal among the others.

3.5. Analytical Figures of Merit

A series of standard solutions containing different MBC concentration was subjected according to the recommended CPE procedures A and B, in order to construct the calibration graphs between the absorbance and MBC concentration as showed in Figures 13 and 14. The statistical data and

analytical figures of merits for MBC in the two proposed CPE-Spectrophotometry are summarized in Table 2. The calibration graphs were linear in the range of 0.5-13 and 1-20 ng mL⁻¹ of MBC with the correlation coefficient of 0.9995 and 0.9997 for method A and B respectively. The percent linearity of 99.89% and 99.94% of the two methods suggest that the calibration curves are statically valid fit. These two fitted linear calibration models were used to estimate MBC concentration in all analysed samples which appear justified, on the statistical basis. The limit of detection and limit of quantitation are calculated using the following formulas; LOD=3σ_B/s; LOQ=10 σ_B/s, where (σ_B) is the standard deviation of the regression line and (s) its slope, and found to be of 0.46 and 0.49 ng mL⁻¹ for the method A and B respectively. The enrichment factor, defined as the ratio of slope of calibration curve obtained by CPE to that obtained without pre-concentration was of 85.7 and 38.9 fold for the method A and B respectively. This in turn enhanced the sensitivity of the spectrophotometric method which was 4.5 times better than that obtained by Naidu et al [24].

Concerning the detection limit, our finding was better than that obtained by other reported methods (Table 3). By considering a limit of detection of 0.46-0.49 μg L⁻¹ in aqueous solution and 15 g of vegetable and fruit samples in 10 mL solution, LOD of the method was also calculated and found in the range 0.0031-0.0032 mg kg⁻¹ for MBC. This finding has encouraged the authors to apply the proposed methods in the estimation of MBC in real samples such as vegetables, fruits and environmental samples to test its applicability and reliability. In fact, the developed methods comply with the requirements of the international standards in terms of the maximum residue limits (MRL) of MBC insecticide in different types of foods set by FAO/WHO and other international bodies [7-8].

3.6. Accuracy and Precision

The accuracy and precision are the most crucial and basic requirements in method validation for ensuring quality and reliability of the results in the applicability of the analytical method. Thus, the accuracy of the proposed methods was examined in terms of percent recovery by the spiking river and soil samples with 1.0, 5.0 and 11.0 ng mL⁻¹ standard MBC for method A and 2.0, 8.0 and 16.0 ng mL⁻¹ standards BDC for method B, from which subjected to the recommended CPE procedures. The results are presented in Tables 4 and 5. It can be seen that a good accuracy in terms of percent recoveries obtained were within average of 97.86±1.06% for Method A and 98.66±0.93% for method B. This confirmed that the systematic errors are relatively absent, concluding the presence of matrix components of these samples have no appreciable effect on the determination of the target analyte. Also, each spiked sample was repeated five times for precision testing in terms of repeatability and found in the range of 1.25-2.97% -for method A and 0.37-1.42% for method B, indicative of a good precise for the proposed method. So, these analytical procedures may be very useful and suitable for the application in the routine environmental and food laboratories.

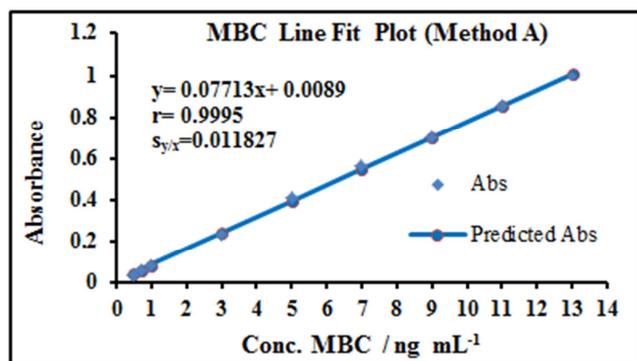


Figure 13. Calibration curve of MBC pesticide for the proposed method A.

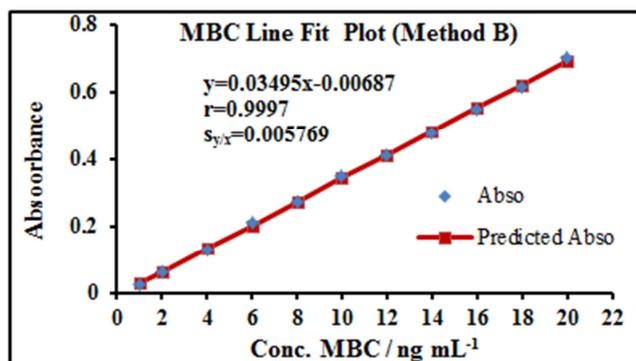


Figure 14. Calibration curve of MBC pesticide for the proposed method B.

Table 1. Effect of order of additions.

| Method | No. | Addition | Abs |
|--------|-----|---|-------|
| A | 1 | MBC+H ₃ PO ₄ +K ₃ Fe(CN) ₆ +FeCl ₃ +Buffer(4)+TX-114 | 0.329 |
| | 2 | MBC+FeCl ₃ +K ₃ Fe(CN) ₆ +Buffer(4)+H ₃ PO ₄ +TX-114 | 0.412 |
| | 3 | MBC+K ₃ Fe(CN) ₆ +FeCl ₃ +Buffer(4)+H ₃ PO ₄ +TX-114 | 0.368 |
| | 4 | MBC+Buffer(4)+FeCl ₃ +K ₃ Fe(CN) ₆ +H ₃ PO ₄ +TX-114 | 0.298 |
| B | 1 | MBC+FeCl ₃ +2,2-Bipyridyl+Buffer(3)+TX-114 | 0.241 |
| | 2 | MBC+Buffer(3)+FeCl ₃ +2,2-Bipyridyl+TX-114 | 0.311 |
| | 3 | MBC+2,2-Bipyridyl+FeCl ₃ +Buffer(3)+Tritonx-114 | 0.410 |
| | 4 | MBC+Buffer(3)+FeCl ₃ +2,2-Bipyridyl+TX-114 | 0.193 |

Table 2. The statistical data and analytical figures of merits for carbendazim by CPE- Spectrophotometry.

| Parameter | Method A | Method B |
|---|-------------------------------|--------------------------------|
| color | Blue | Orange |
| λ_{max} (nm) | 685 | 521 |
| Regression equation | $y = 0.07713x + 0.0089$ | $y = 0.02395x - 0.00687$ |
| Standard deviation of regression line ($S_{y/x}$) | 0.011827 | 0.005769 |
| Correlation coefficient (r) | 0.9995 | 0.9997 |
| Linearity percentage (%R) | 99.89 | 99.94 |
| C.L. for the slope ($b \pm ts_b$) at 95% | 0.07713 ± 0.0021 | 0.02395 ± 0.00063 |
| C.L. for the intercept ($a \pm ts_a$) at 95% | 0.0089 ± 0.01501 | -0.00687 ± 0.00753 |
| Beer's law range (ng mL ⁻¹) | 0.5-13 | 1-20 |
| Limit of Detection (ng mL ⁻¹) | 0.46 | 0.49 |
| Limit of Quantitation (ng mL ⁻¹) | 1.53 | 1.65 |
| Sandell's sensitivity ($\mu\text{g cm}^{-2}$) | 0.013 | 0.029 |
| Molar absorptivity (L.mol ⁻¹ .cm ⁻¹) | 2.07×10^4 | 1.83×10^4 |
| RSD% (n=5) | 2.15 at 1 ng mL ⁻¹ | 3.25 at 4 ng mL ⁻¹ |
| RSD% (n=5) | 1.64 at 9 ng mL ⁻¹ | 2.74 at 16 ng mL ⁻¹ |
| Preconcentration factor* | 50 | 58.8 |
| Enrichment factor | 85.7 | 38.9 |
| Extraction efficiency(%E)** | 98.2 | 98.0 |

*Preconcentration factor was calculated the ratio of the original sample volume to that of extracted volume (of surfactant-rich phase)** Extraction efficiency was calculated according to the following formula,

$$\%E = [(1 - C_w)(Rv + 1)CSR] \times 100$$

where R_v is the volume ratio of surfactant-rich phase (SRP) to the aqueous phase. C_w is the concentration of analyte in aqueous phase (original solution before CPE), and C_{SRP} is the concentration of target analyte in SRP, which was quantified using calibration curve obtained from the original solutions (without CPE).

Table 3. Comparison of the proposed methods with other reported methods for the determination of MBC.

| Extraction Procedure | Detection System | Linear Range (ng mL ⁻¹) | Limit of Detection (ng mL ⁻¹) | Ref. |
|----------------------|------------------|-------------------------------------|---|------|
| SPE | HPLC | 25-500 | 3.55 | [27] |
| SPE | HPLC | - | 20 | [29] |
| SPME | HPLC | 10-1000 | 1.0 | [30] |
| DLLME | HPLC | 5-800 | 0.5 | [31] |
| DLLME | UV-Vis | 5-600 | 2.1 | [32] |
| IL-DLLME | HPLC | 5-500 | 5 | [33] |
| CPE | UV-Vis | 0.5-13 | 0.46 | This |
| | | 1.0-20 | 0.49 | work |

Table 4. The accuracy and precision of the proposed method (A) for the determination of MBC by CPE-Spectrophotometry.

| sample | Amount MBC taken (ng mL ⁻¹) | Amount MBC found (ng mL ⁻¹) | Rec (%) | Ave. Recovery (%) (x±ts/√n) | E _{rel} (%) | %RSD (n=5) |
|-------------|---|---|---------|-----------------------------|----------------------|------------|
| River Water | 1 | 0.98 | 98.00 | 97.86±1.06% | -2.00 | 2.31 |
| | 5 | 4.88 | 97.60 | | -2.40 | 1.93 |
| | 11 | 10.85 | 98.64 | | -1.36 | 1.25 |
| Soil | 1 | 0.96 | 96.0 | | -4.0 | 2.97 |
| | 5 | 4.94 | 98.8 | | -1.2 | 2.10 |
| | 11 | 10.79 | 98.1 | | -1.9 | 1.77 |

Table 5. The accuracy and precision of the proposed method (B) for the determination of MBC by CPE-Spectrophotometry.

| sample | Amount MBC taken (ng mL ⁻¹) | Amount MBC found (ng mL ⁻¹) | Rec (%) | Ave. Recovery (%) at 95% C.I (x±ts/√n) | E _{rel} (%) | %RSD (n=5) |
|-------------|---|---|---------|--|----------------------|------------|
| River Water | 2 | 1.98 | 99.00 | 98.66±0.93% | -1.00 | 1.42 |
| | 8 | 7.95 | 99.38 | | -0.62 | 1.09 |
| | 16 | 15.89 | 99.32 | | -0.68 | 0.73 |
| Soil | 2 | 1.94 | 97.00 | | -3.0 | 1.01 |
| | 8 | 7.87 | 98.38 | | -1.62 | 0.86 |
| | 16 | 15.82 | 98.88 | | -1.12 | 0.37 |

3.7. Interferences Study

The study was conducted by addition of various amounts of expected interfering species in the samples under study to the standard solution containing 10 ng mL⁻¹ of MBC followed the general CPE procedure (A) and (B), to verify more whether these interfering species affect the accuracy of the proposed methods. The results are shown in Tables 6 and 7. The results revealed that the metal ions and other compound species at different amounts do not effect on the percent recovery levels of each pesticide, indicating no appreciable interferences exist, affect the determination of MBC pesticide and concluding that good selectivity has achieved for the proposed methods.

3.8. Applications

According to the initial analysis of the study samples, it was shown that there is no existence of any residues of MBC can be detected in these samples by the proposed methods. Therefore, all the selected samples were spiked with MBC standard at concentration level of 3.0, 7.0 and 13.0 ng mL⁻¹ for method A and 4.0, 10.0 and 18.0 ng mL⁻¹ for method B, and then subjected to the recommended CPE procedure for five replicates measurements and the target MBC concentration in each spiked sample was measured spectrophotometrically at each respective absorption maximum. The results are summarized in Tables 8 and 9. In all cases, it can be seen that the percent recovery levels for MBC pesticide in all sample were acceptable and ranging from 95.00% to 100.14% with standard deviation from 0.97 to 2.49 for method A and from 96.50% to 100.50% with standard deviation from 0.71 to 2.49 for method B.

The results of the two methods were also subjected to statistical treatment by using t-test at 95% confidence interval to test the significant of the two methods by using t-test: two-sample assuming equal variances as showed in Table 10. It can be seen that the calculated t-value [t] (=1.17285) tested whether at one-tail or two-tail at P (T<=t) =0.05 for 22 degrees of freedom is less than the critical values, so the null

hypothesis (Ho) is retained, concluding there is no evidence to suggest that the method A is significantly difference from Method B in accuracy (p=0.13, p=0.25). Also, F-test has shown that the calculated F-value (F=1.288) whether at one-tail or two-tail at P(T<=t)=0.05 for 11 degrees of freedom is less than the critical values (F_{0.05,11,11}=2.81 one-tail), so the null hypothesis (Ho) is accepted, concluding that there appears no statistical evidence to suggest that the variability of the two methods is significantly different in the precision (p=0.34).

Table 6. Effect of diver's species on the percent recovery of MBC by the proposed method A.

| Foreign species | Recovery % | | | Recovery % mean±SD |
|------------------|------------|----------|----------|--------------------|
| | 250 (ng) | 500 (ng) | 750 (ng) | |
| K ⁺ | 99.84 | 100.01 | 100.5 | 100.11±0.34 |
| Ca ²⁺ | 99.26 | 99.14 | 100.78 | 99.72±0.91 |
| Mg ²⁺ | 99.57 | 100.90 | 99.40 | 99.95±0.82 |
| Fe ²⁺ | 98.94 | 100.57 | 101.12 | 100.21±1.13 |
| Co ²⁺ | 98.37 | 99.03 | 100.01 | 99.13±0.82 |
| Vitamin B | 98.80 | 99.60 | 99.90 | 99.43±0.56 |
| Vitamin C | 96.98 | 96.40 | 95.32 | 96.23±0.84 |
| Glucose | 99.58 | 100.04 | 101.09 | 100.23±0.77 |
| Fructose | 98.35 | 99.74 | 100.07 | 99.38±0.91 |
| Protein | 97.00 | 99.71 | 99.96 | 98.89±1.64 |

Table 7. Effect of diver's species on the percent recovery of MBC by the proposed method B.

| Foreign species | Recovery % | | | Recovery % mean±SD |
|------------------|------------|--------|--------|--------------------|
| | 250 ng | 500 ng | 750 ng | |
| K ⁺ | 99.83 | 99.12 | 98.50 | 99.15±0.66 |
| Ca ²⁺ | 98.28 | 99.57 | 100.04 | 99.29±0.91 |
| Mg ²⁺ | 97.85 | 99.39 | 100.50 | 99.24±1.33 |
| Fe ²⁺ | 99.64 | 98.61 | 100.97 | 99.74±1.18 |
| Co ²⁺ | 99.25 | 96.93 | 98.71 | 98.29±1.21 |
| Vitamin B | 98.83 | 96.69 | 95.90 | 97.14±1.51 |
| Vitamin C | 97.38 | 97.40 | 95.32 | 96.70±1.19 |
| Glucose | 99.87 | 97.52 | 100.06 | 99.15±1.41 |
| Fructose | 97.54 | 99.17 | 96.74 | 97.81±1.23 |
| Protein | 99.00 | 97.33 | 96.95 | 97.66±1.09 |

Table 8. Analytical results of carbendazim in different samples by proposed method A.

| Sample | MBC added (ngmL ⁻¹) | MBC found (ng mL ⁻¹ ±SD) | Recovery% (mean±SD) | RSD% n=5 |
|-----------|---------------------------------|-------------------------------------|---------------------|----------|
| Tap water | 3 | 2.98±0.04 | 99.33±1.24 | 1.39 |
| | 7 | 6.97±0.08 | 99.57±0.97 | 1.16 |
| | 13 | 12.86±0.10 | 98.92±0.71 | 0.83 |
| Cucumber | 3 | 2.85±0.07 | 95.00±2.49 | 2.63 |
| | 7 | 7.01±0.12 | 100.14±1.74 | 1.84 |
| | 13 | 12.76±0.15 | 98.15±1.22 | 1.25 |
| Tomato | 3 | 2.93±0.06 | 97.66±2.04 | 2.09 |
| | 7 | 6.84±0.12 | 97.71±1.70 | 1.74 |
| | 13 | 12.79±0.13 | 98.38±0.97 | 0.99 |
| Orange | 3 | 2.94±0.05 | 98.00±1.79 | 1.83 |
| | 7 | 6.90±0.10 | 98.57±1.45 | 1.48 |
| | 13 | 12.87±0.15 | 99.00±1.13 | 1.15 |

Table 9. Analytical results of carbendazim in different samples by proposed method B.

| Sample | MBC added (ngmL ⁻¹) | MBC found (ng mL ⁻¹ ±SD) | Recovery% (mean±SD) | RSD% n=5 |
|-----------|---------------------------------|-------------------------------------|---------------------|----------|
| Tap water | 4 | 3.93±0.05 | 98.25±1.24 | 1.31 |
| | 10 | 9.92±0.09 | 99.20±0.97 | 0.95 |
| | 18 | 17.87±0.15 | 99.27±0.71 | 0.83 |
| Cucumber | 4 | 3.86±0.08 | 96.50±2.49 | 2.13 |
| | 10 | 9.84±0.17 | 98.40±1.74 | 1.72 |
| | 18 | 18.01±0.19 | 100.05±1.22 | 1.11 |
| Tomato | 4 | 3.90±0.09 | 97.50±2.04 | 2.23 |
| | 10 | 9.92±0.16 | 99.20±1.70 | 1.63 |
| | 18 | 17.95±0.16 | 99.72±0.97 | 0.91 |
| Orange | 4 | 3.99±0.06 | 99.75±1.79 | 1.55 |
| | 10 | 10.05±0.12 | 100.50±1.45 | 1.18 |
| | 18 | 17.84±0.15 | 99.11±1.13 | 0.82 |

Table 10. Statistical comparison between Method A and B for the determination of MBC in real samples by CPE- spectrophotometry.

| Method A | Method B | Statistical Parameters | Method A | Method B |
|----------|----------|----------------------------|----------|----------|
| 99.33 | 98.25 | Mean | 98.38583 | 98.97083 |
| 99.57 | 99.20 | Variance | 1.680808 | 1.304608 |
| 98.92 | 99.27 | Observations | 12 | 12 |
| 95.00 | 96.50 | Pooled Variance | 1.492708 | |
| 100.14 | 98.40 | dof | 22 | |
| 98.15 | 100.05 | t Stat | -1.17285 | |
| 97.66 | 97.50 | P(T<=t) one-tail | 0.126699 | |
| 97.91 | 99.20 | t Critical one-tail | 1.717144 | |
| 98.38 | 99.72 | P(T<=t) two-tail | 0.253398 | |
| 98.00 | 99.95 | t Critical two-tail | 2.073873 | |
| 98.57 | 100.5 | F Stat | 1.288 | |
| 99.00 | 99.11 | F _{crit} , dof=11 | 2.81 | |

4. Conclusion

This work accentuates the established methods based on designed two CPE procedures before spectrophotometric

analysis for the determination of ultra trace carbendazim residues in a variety of samples. Despite various analytical methods reported in a chemical literature for the detection of MBC in different samples, to our knowledge this is the first CPE-spectrophotometry for MBC depended on the reaction of the target analyte with the well-known reaction system reported in a scientific literature. The proposed methods provide good analytical features such as a linear range, limit of detection, accuracy, precision and interferences-free, which can be applicable for routine quantitative analysis in the quality control laboratories for MBC in the environmental samples, due to the simplicity, rapidity, inexpensive and eco-friendly.

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References

- Venetian A, Vacca, G, Arana S, De Simone F and L. Rastrelli (2004) Determination of carbendazim, thiabendazole and thiophanate-methyl in banana (*Musa acuminata*) samples imported to Italy. Food Chem 87: 383-386.
- Tomlin C Ed. (1994) The Pesticide Manual, 10th ed. British Crop Protection Council and Royal Society of Chemistry, U.K., p. 149-50.
- Mazo LH, Coutinho CF, Galli A. and Machado SAS (2006) Carbendazim EO meio ambiente: degradação e toxidez. Pesticidas (UFPR), v. 16, p. 63-70.
- Olayemi OA (2015). Comparative toxicity of two different pesticides on the skin of Japanese quail (*Cortunix japonica*). World Vet J 5: 13-18.
- European Communities; Council Directive 98/83/EC on the Quality of Water Intended for Human Consumption L 330/32; Official Journal of the European Communities: Brussels, December (1998).
- United States Environmental Protection Agency (US EPA); Drinking Water Contaminants; U.S. Agency for International Development: Washington D.C. May (2009).
- Codex Alimentarius Commission. Pesticides residue in food. Joint FAO/ WHO Food Standard Programme of United Nations. Vol. 2, 2nd ed. Rome. (1993).
- GB 14870-1994: Maximum Residue Limits of Carbendazim in Foods; Standardization Administration of China (SAC), General Administration of Quality Supervision, Inspection and Quarantine of the People's Republic of China: Beijing (1994).
- ANVISA, Agencia Nacional de Vigilancia Sanitaria, Public Consultation No. 113 of December 19th, (2007).
- Bushway RJ, Hurst HL, Kagabalasooriar J, and Perkins LB (1991) Determination of carbendazim in blueberries by reversed phase high-performance liquid chromatography. J Chromatogr 587: 321-24.

- [11] Regis-rolle SD and Bauville GM (1993) High-performance liquid chromatographic method for the determination of carbendazim residues in crops, grains, and wines with fluorescent detection. *Pestic Sci* 37: 273-82.
- [12] Phansawan B, Prapamonto T, Thavornnyutikarn P, Chantara S, Mangklabruks A and C. Santasup C (2015) A Sensitive method for determination of carbendazim residue in vegetable samples using HPLC-UV and its application in health risk assessment. *Chiang Mai J Sci* 42: 681-690.
- [13] Thomas DH, Lopez-Avila V, Betowski LD and Van Emon J (1996) Determination of carbendazim in water by high-performance immunoaffinity chromatography on-line with high-performance liquid chromatography with diode-array or mass spectrometric detection. *J Chromatogr A* 724: 207-217.
- [14] Vega BA, Frenich GA and Vidal MLJ (2005) Monitoring of pesticides in agricultural water, soil samples from Andalusia by liquid chromatography coupled to mass spectrometry. *Anal Chim. Acta* 538: 117-127.
- [15] Cheng P, Tang HM, Yang SX and L. Wang L (2009) Determination of carbendazim residue in tomato and cucumber by HPLC-MS-MS. *Modern Agrochemicals* 3: 36-37.
- [16] Steinwandter H (1985) Chemical derivatization in residue analysis in gas chromatographic determination of carbendazim after alkylation with diazomethane and diazoethane. *Fresenius' Z Anal Chem* 321: 599-600.
- [17] Subhani Q, Huang Z, Zhu Z. and Zhu Y (2013) Simultaneous determination of imidacloprid and carbendazim in water samples by ion chromatography with fluorescence detector and post-column photochemical reactor. *Talanta* 116: 127-132.
- [18] Ashrafi AM, Đorđević J, Guzsvány V, Švancara I, Trtić-Petrović T, Purenović M and Vytřas K (2012) Trace determination of carbendazim fungicide using adsorptive stripping voltammetry with a carbon paste electrode containing tricresyl phosphate. *Int J Electrochem Sci* 7: 9717-9731.
- [19] Manisankar P, Selvanathan G and Vedhió C (2005) Utilisation of polypyrrole modified electrode for the determination of pesticides. *Int J Environ Anal Chem* 85: 409-422.
- [20] Li J and Chi Y (2009) Determination of carbendazim with multiwalled carbon nanotubes-polymeric methyl red film modified electrode. *Pesticide Biochem Physiol* 93: 101-104.
- [21] Ribeiro WF, Selva TMG, Lopes IC, Coelho ECS, Lemos SG, de Abreu FC, do Nascimento VB and de Araujo MCU (2011) Electroanalytical determination of carbendazim by square wave adsorptive stripping voltammetry with a multiwalled carbon nanotubes modified electrode. *Anal Methods* 3: 1202-1206.
- [22] Itak JA, Selisker MY, Jourdan SW, Fleeker JR and Herzogt DP (1993) Determination of benomyl (as carbendazim) and carbendazim in water, soil, and fruit juice by a magnetic particle-based immunoassay. *J Agric Food Chem* 41: 2329-2332.
- [23] Zhu HS, Wu LH, Li RB, Xia LA, Han JQ, Zhang QJ, Bian YC and Yu QR (2008) Determination of pesticides in honey using excitation – emission matrix fluorescence coupled with second – order calibration and second – order addition methods. *Anal Chim Acta* 619: 165-172.
- [24] Naidu KP, Niranjan T and Naidu VS (2011) Spectrophotometric determination of carbendazim in its formulations and environmental samples. *Intern J ChemTech Res* 3: 1728-1733.
- [25] Wu YS and Lee HK (1997) Determination of carbendazim residues in grains by solid-phase extraction and micellar electrokinetic chromatography with ultraviolet detection. *J Chromatogr Sci* 35: 513-518.
- [26] Ebaisat H (2011) Determination of some benzimidazole fungicides in tomato puree by high performance liquid chromatography with SampliQ polymer SCX solid phase extraction. *Arab J Chem* 4: 115-117.
- [27] Kong HX, Yun H and Qiu NX (2007) Determination of carbendazim residue in apple juice concentrate by high performance liquid chromatography with solid-phase extraction. *Chin J Anal Lab* 26: 65-67.
- [28] Michel M and Buszewski B (2004) Optimization of a matrix solid-phase dispersion method for the determination analysis of carbendazim residue in plant material. *J Chromatogr B* 800: 309-314.
- [29] Hu Y, Yang X, Wang Z, Wang C and Zhao J (2005) Determination of carbendazim and thiabendazole in tomatoes by solid-phase microextraction coupled with high performance liquid chromatography and fluorescence detection. *Chin J Chromatogr* 23: 581-584.
- [30] Wu Q, Li Y, Wang C, Liu Z, Zang Z, Zhou X and Wang Z (2009) Dispersive liquid liquid microextraction combined with high performance liquid chromatography-fluorescence detection for the determination of carbendazim and thiabendazole in environmental samples. *Anal Chim Acta* 638: 139-145.
- [31] Pourreza N, Rastegarzadeh S and Larki A (2015) Determination of fungicide carbendazim in water and soil samples using dispersive liquid-liquid microextraction and microvolume UV-Vis spectrometry. *Talanta* 134: 24-29.
- [32] Asensio-Ramos M, Hernández-Borges J, Borges-Miquel, TM and Rodríguez-Delgado A (2011) Ionic liquid-dispersive liquid-liquid microextraction for the simultaneous determination of pesticides and metabolites in soils using high-performance liquid chromatography and fluorescence detection. *J Chromatogr A* 1218: 4808-4816.
- [33] Zhou ZM, Chen JB, Zhao DY and Yang MM (2009) Determination of four carbamate pesticides in corn by Cloud point extraction and high-performance liquid chromatography in the visible region based on their derivatization reaction. *J Agr Food Chem* 57: 8722-8727.
- [34] Tang T, Qian K, Shi T, Wang F, Li J, and Cao Y (2010) Determination of triazolefungicides in environmental water samples by high performance liquid chromatography with cloud point extraction using polyethylene glycol 600 monooleate. *Anal Chim Acta* 80: 26-31.
- [35] Chen JB, Zhao WJ, Liu W, Zhou ZM and Yang MM (2009) Cloud point extraction coupled with derivative of carbofuran as a preconcentration step prior to HPLC. *Food Chem* 115: 1038-1041.
- [36] Melchert WR and Rocha FRP (2009) Cloud point extraction and concentration of carbaryl from natural waters. *Intern J Environ Anal Chem* 89: 969-979.
- [37] Khammas ZAA and Ahmad SS. (2016) Micelle-mediated extraction combined with visible spectrophotometry for the determination of ultra trace amounts of bendiocarb insecticide in various matrices after oxidative coupling with O-Toluidine. *Int Res J Pure Appl Chem* 10: 1-16.

- [38] Schenck FJ and Hobbs JE (2004) Evaluation of the quick, easy, cheap, effective, rugged, and safe (QuEChERS) approach to pesticide residue analysis. *Bull Environ Contamint and Toxicol* 73: 24–30.
- [39] Theis TL and Singer PC (1973) the stabilization of ferrous iron by organic compounds in natural water, In P. C. Singer [ed.], *Trace metals and metal-organic interactions in natural waters*. Ann Arbor Sci. p. 303-320.
- [40] Theis TL and Singer PC (1973) Complexation of iron (II) by organic matter and its effect on iron (II) oxygenation. *Enviro. Sci Technol* 8: 569-573.
- [41] Stumm W and Lee GF. (1960) The chemistry of aqueous iron. *Schweiz Z Hydrol* 22: 295-319.
- [42] Mohammed AA, Talaat E, Mohamed, Y and Shama SA (2012) Application of oxidants to the spectrophotometric microdetermination of meclizine HCl in pure and pharmaceutical formulations. *Prime J Microbiol Res* 2: 137-140.